

THE EFFECT OF ELEVATED TEMPERATURE ON COPPER TOXICITY TO THE THERMOPHILIC ALGA *SYNECHOCOCCUS LIVIDUS* (CYANOPHYCEAE)

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ABSTRACT.—The hypothesis that temperature influences the toxicity of copper to thermophilic Cyanophyceae was tested in a laboratory study with *Synechococcus lividus*. This thermophile was tested at copper concentrations from 0 to 200 $\mu\text{g/l}$, and temperatures from 40.0 to 50.0 C. It was found that an interaction between increased copper and temperature significantly decreased the rate of carbon assimilation, chlorophyll content, and photosynthetic efficiency.

Geothermal springs represent a remarkably unique and stable environment with respect to a large number of physical and chemical parameters. The temperatures of these springs rarely vary more than 2 C throughout the seasons. A constant flow rate with laminar flow characteristics exists which, among other things, minimizes the forces of erosion. Light intensity is high. The area around hot springs is usually devoid of trees and the water column is shallow. Nutrient replenishment is continuous in the flowing water system, so that nutrient deficiencies probably do not develop (Brock 1970).

Even in these seemingly ideal conditions, only a restricted flora exists. Due to the elevated temperatures of the thermal spring environment—approximately 50 C to well above the boiling point in fumaroles—prokaryotes are usually the sole inhabitants (Brock 1967a). The Cyanophyceae present have an upper temperature limit of 73–75 C. These algae are not merely subsisting, but are actually growing and thriving at a given location (Brock 1967b). This heat tolerance seems to be due to a number of factors, including the thermal stability of their photosynthetic membrane systems, the low Q_{10} value of respiratory rates preventing acceleration to lethal catabolism, the heat stability of the algal protoplasmic structures and the capacity of their proteins to endure high temperatures without denaturation, and the lack of competition in the environment

(Brock 1974, Lewin 1962). Luxuriant growth is to be expected in these locations.

However, visible degradation of the algal mats has occurred in many thermal springs of Yellowstone National Park. This deterioration is particularly noticeable in those areas which are heavily frequented by visitors.

Changes in water temperature, nutrient concentration, flow rate, etc., may be eliminated as possible mechanisms for degradation due to the stability of the environment. An external factor exists as the remaining possibility—i.e., the introduction of copper coinage to the thermal springs (R. A. Hutchinson, Yellowstone National Park Geologist, pers. comm.).

Copper has long been widely used as an algicide. The recommended dose for algal control in alkaline water ranges from 0.2 to 2.0 mg/l (Trainor 1978), but can be as low as 50 $\mu\text{g/l}$ for *Chlorella* (Bartlett et al. 1974). The chemical analysis of some of the major thermal springs of Yellowstone National Park indicates the copper concentration ranges from 1 to 9 $\mu\text{g/l}$ (Brock, 1978), which appears to be much too low for the demonstrated algicidal effects. However, this does not preclude the possibility of increased toxicity at the elevated temperatures found in the thermal springs. These temperatures approach the critical maximum for life itself.

The possibility, therefore, exists that even low copper concentration in a thermal environment produce a detrimental effect on

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algal mats—i.e., an interactive effect exists between copper and heat.

METHODS

Axenic cultures of *Synecococcus lividus* (R. Castenholz, Department of Biology, University of Oregon, pers. comm.) were maintained in a general growth medium (Miller et al. 1978) with a 12-hour light:dark cycle. Cultures were frequently diluted to maintain cells in exponential growth phase.

Batch cultures were acclimated to a temperature (± 0.01 C) for one week prior to each experiment. Following this acclimation period the culture volume was subdivided, and to each aliquot an amount of copper was added as $\text{CuSO}_4 \cdot 5\text{H}_2\text{O}$. Final concentrations

were 0, 50, 100, 150, and 200 $\mu\text{g Cu/l}$. Aliquots were then incubated a further 24 hours after which each treatment was dispensed into triplicate 125 ml glass bottles for measurement of carbon assimilation, and an amount was filtered for chlorophyll analysis.

To each of the replicate bottles, we added 5 μCi of ^{14}C bicarbonate (New England Nuclear). Cells were incubated for three hours, then membrane filtered and washed with distilled water. Filters were dried, placed in omnifluor, and activity measured by liquid scintillation. Because each treatment was handled in the same manner, radioactive counts per minute (CPM) were directly comparable among treatments. Chlorophyll-a was estimated from the optical density of ethanol extracts.

TABLE 1. Relative carbon assimilation (counts per minute) at experimental temperatures and copper concentrations.

| Degrees C | $\mu\text{g/l Cu}$ | | | | |
|-----------|--------------------|--------|-------|-------|------|
| | 0 | 50 | 100 | 150 | 200 |
| 40.0 | 31409 | 8339 | 8595 | 7505 | 2922 |
| | 47990 | 10980 | 8027 | 4508 | 1259 |
| | 49364 | 9436 | 7765 | 4059 | 1032 |
| | 46400 | 11315 | 11415 | 3382 | 2662 |
| | 44649 | 10932 | 9174 | 2421 | 1284 |
| | 44459 | 9157 | 11700 | 4087 | 1221 |
| 42.5 | 28340 | 26293 | 9845 | 2294 | 1076 |
| | 31502 | 16756 | 8746 | 2468 | 631 |
| | 39546 | 21045 | 7263 | 1845 | 463 |
| | 45936 | 13467 | 8034 | 2932 | 1399 |
| | 51584 | 12640 | 6336 | 1317 | 2477 |
| | 46953 | 12162 | 7490 | 3420 | 352 |
| 45.0 | 218047 | 131760 | 53929 | 26266 | 8420 |
| | 195315 | 138318 | 53482 | 24028 | 6977 |
| | 165850 | 126580 | 46425 | 24480 | 4832 |
| | 184091 | 113685 | 39251 | 23552 | 2999 |
| | 178200 | 131783 | 29053 | 21473 | 2426 |
| | 147796 | 137110 | 38914 | 20091 | 1499 |
| 47.5 | 126228 | 46272 | 16153 | 4162 | 1033 |
| | 140663 | 29294 | 13862 | 2798 | 903 |
| | 155604 | 26466 | 10830 | 2287 | 1459 |
| | 164761 | 22688 | 11420 | 1932 | 619 |
| | 155644 | 26186 | 11849 | 1257 | 580 |
| | 147319 | 22365 | 10548 | 2040 | 520 |
| 50.0 | 53933 | 37702 | 11616 | 4812 | 521 |
| | 54506 | 33803 | 11605 | 6573 | 1762 |
| | 62672 | 24093 | 12541 | 1846 | 803 |
| | 73887 | 21548 | 9130 | 1942 | 4422 |
| | 79563 | 17767 | 7518 | 2232 | 506 |
| | 62200 | 17893 | 12252 | 4387 | 826 |

This procedure was repeated at each of five growth temperatures: 40.0, 42.5, 45.0, 47.5, and 50.0 C. To facilitate statistical manipulation, all CPM and chlorophyll data were normalized with respect to the control treatment (0 $\mu\text{g/l}$ Cu) to eliminate differences among treatments due to variations in starting population density.

RESULTS AND DISCUSSION

Carbon assimilation data in terms of CPM are shown in Table 1. At all experimental temperatures increased copper concentration led to decreased carbon assimilation so that, at 200 μg , Cu/l assimilation was less than 5 percent of the control value. Using analysis of variance for two-way classification (Mendenhall et al. 1977) we found a significant interaction ($F_{16,125} = 3.31$, $p < 0.01$) between temperature and copper concentration.

Two possible causes for decreased carbon assimilation include decreased chlorophyll content per cell and depressed photosynthetic efficiency measured as carbon assimilation per unit chlorophyll. Table 2 shows

the chlorophyll concentration of aliquot cultures after only 24 hours of incubation in the presence of copper. Again, at all experimental temperatures there is a significant decrease in chlorophyll with increased copper. There was also significant interaction ($F_{16,25} = 8.56$, $p < 0.01$) between temperature and copper.

Photosynthetic efficiency as measured by carbon assimilation per unit chlorophyll similarly decreased with increasing copper concentration (Table 3). Further, there was a significant interaction ($F_{16,125} = 12.44$, $p < 0.01$) between temperature and copper.

From these data it appears that copper interacting with temperature can cause significant depression of the photosynthetic activity of *Synechococcus lividus*. This appears to be caused by a decrease in chlorophyll content of the cell, and a lowered photosynthetic efficiency. It is possible that at temperatures higher than those examined only minute concentrations of copper may prove to be toxic to *S. lividus*. If this is so, and if our results may be extended to other thermophilic cyanophytes, this is a possible mechanism to

TABLE 2. Chlorophyll-a concentrations ($\mu\text{g/l}$) of aliquot cultures following 24-hour incubation with various copper concentrations.

| Degrees C | $\mu\text{g/l}$ Cu | | | | |
|-----------|--------------------|----|-----|-----|-----|
| | 0 | 50 | 100 | 150 | 200 |
| 40.0 | 52 | 42 | 41 | 37 | 31 |
| | 49 | 40 | 37 | 36 | 29 |
| 42.5 | 50 | 38 | 27 | 31 | 28 |
| | 47 | 36 | 34 | 31 | 28 |
| 45.0 | 98 | 80 | 64 | 75 | 58 |
| | 98 | 69 | 64 | 77 | 61 |
| 47.5 | 87 | 63 | 59 | 42 | 34 |
| | 90 | 62 | 66 | 41 | 33 |
| 50.0 | 85 | 55 | 48 | 37 | 27 |
| | 83 | 58 | 51 | 38 | 26 |

TABLE 3. Relative photosynthetic efficiency as CPM ^{14}C assimilated per mg chlorophyll. Control treatments were normalized to 100 percent.

| Degrees C | $\mu\text{g/l}$ Cu | | | | |
|-----------|--------------------|----|-----|-----|-----|
| | 0 | 50 | 100 | 150 | 200 |
| 40.0 | 100 | 28 | 28 | 14 | 6 |
| 42.5 | 100 | 55 | 31 | 9 | 5 |
| 45.0 | 100 | 94 | 37 | 16 | 4 |
| 47.5 | 100 | 27 | 12 | 3 | 2 |
| 50.0 | 100 | 59 | 28 | 12 | 7 |

explain the current deterioration of algal mats in many thermal springs.

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